# Metallic Iron for Environmental Remediation: Learning from the Becher

### **Process**

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#### 7 Abstract

- Metallic iron (Fe<sup>0</sup>) is a moderately reducing agent that has been reported to be capable of 8 9 reducing many environmental contaminants. Reduction by Fe<sup>0</sup> used for environmental 10 remediation is a well-known process to organic chemists, corrosion scientists and hydrometallurgists. However, considering Fe<sup>0</sup> as a reducing agent for contaminants has faced 11 considerable scepticism because of the universal role of oxide layers on Fe<sup>0</sup> in the process of 12 electron transfer at the Fe<sup>0</sup>/oxide/water interface. This communication shows how progress 13 14 achieved in developing the Becher process in hydrometallurgy could accelerate the comprehension of processes in Fe<sup>0</sup>/H<sub>2</sub>O systems for environmental remediation. The Becher 15 16 process is an industrial process for the manufacture of synthetic rutile (TiO<sub>2</sub>) by selectively removing metallic iron (Fe<sup>0</sup>) from reduced ilmenite (RI). This process involves an aqueous 17 18 oxygen leaching step at near neutral pH. Oxygen leaching suffers from serious limitations 19 imposed by limited mass transport rates of dissolved oxygen across the matrix of iron oxides from initial Fe<sup>0</sup> oxidation. In a Fe<sup>0</sup>/H<sub>2</sub>O system pre-formed oxide layers similarly act as 20 physical barrier limiting the transport of dissolved species (including contaminants and O<sub>2</sub>) to 21 the Fe<sup>0</sup>/H<sub>2</sub>O interface. Instead of this universal role of oxide layers on Fe<sup>0</sup>, improper 22 23 conceptual models have been developed to rationalize electron transfer mechanisms at the Fe<sup>0</sup>/oxide/water interface. 24
- 25 **Keywords**: *Becher process*, Leaching, Oxide layers, Remediation, Zerovalent Iron.
- 26 **Capsule:** Results from *Becher process* are valuable for mass transfer in Fe<sup>0</sup>/H<sub>2</sub>O systems.

#### Introduction

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Permeable reactive barriers of elemental iron (Fe<sup>0</sup> walls or remediation Fe<sup>0</sup>/H<sub>2</sub>O systems) are a valuable technological application that has been shown to be both environmentally friendly and cost-effective in the removal of various substances from contaminated waters [1-7]. Since its development in 1990 by Canadian hydrogeologists numerous papers have been written on the topic and approximately 120 iron walls installed worldwide [4]. Even though the importance of the mechanism of contaminant removal was recognised in the early stage of technology development [1,8], there is still discussion about the mechanism of contaminant removal in Fe<sup>0</sup>/H<sub>2</sub>O systems [9-12]. The main reason cited for reported discrepancies was the improper consideration of the significance of the oxide-film on the process of contaminant removal [11,12]. In fact, after considerable initial scepticism [2], it was accepted that contaminants are removed in Fe<sup>0</sup>/H<sub>2</sub>O systems by an abiotic reductive transformation. Based on this consensus, "the role of oxides in reduction reactions" were discussed [13]. Accordingly, the oxide layers may act either as: (i) a semiconductor to relay electron transfer from Fe<sup>0</sup> to the contaminants (direct reduction) or (ii) Fe<sup>II</sup> adsorbent to yield very reactive adsorbed Fe<sup>II</sup> (so-called structural Fe<sup>II</sup> for indirect reduction). Additionally, electron transfer from the Fe<sup>0</sup> body exposed by pitting of the oxide layer to contaminants is likely to occur. The objective of this paper is to show how the consideration of progress achieved in developing the *Becher process* in the hydrometallurgy, could accelerate the comprehension of processes in Fe<sup>0</sup> walls by carefully surveying the impact of the oxide-layer on contaminant mass transfer to the Fe<sup>0</sup>/H<sub>2</sub>O interface. For the sake of clarity the *Becher process* is presented first.

### The Becher process

The *Becher process* is an environmentally friendly, cost-effective extraction method for upgrading ilmenite from 55 % TiO<sub>2</sub> to about 94 % TiO<sub>2</sub> by reduction with charcoal [14-22]. A typical feed made up of 55 % ilmenite (FeTiO<sub>3</sub>), 38 % pseudorutile (Fe<sub>2</sub>Ti<sub>3</sub>O<sub>9</sub>) and 7 % rutile (TiO<sub>2</sub>) is heated in air at about 1000 °C to oxidise the Fe<sup>II</sup> in the ilmenite to Fe<sup>III</sup> resulting in a

phase known as pseudobrookite ( $Fe_2TiO_5 = Fe_2O_3 + TiO_2$ ). Pseudobrookite is then reduced with coal (C/CO) in an iron reduction kiln at 1100 °C. This carbothermic reaction reduces the iron in the pseudobrookite to the metallic (Fe<sup>0</sup>). The reduction product (reduced rutile - RI) consists mainly of grains of Fe<sup>0</sup> embedded in a matrix of titanium dioxide [19,22]. After cooling RI (approximately 2 Fe + TiO<sub>2</sub>), embedded Fe<sup>0</sup> should be leached from (or "rusted out" of) the matrix (about 58 % Fe) at near neutral pH. Leaching Fe<sup>0</sup> from RI is achieved by oxidation with dissolved oxygen from air (so-called "aerial oxidation") mostly with ammonium chloride (NH<sub>4</sub>Cl) acting as a catalyst for rusting. Since the solution is not buffered, the pH will rise resulting in the formation of various iron oxides. Magnetite (Fe<sub>3</sub>O<sub>4</sub>) is the preferred product since it can be magnetically separated from TiO<sub>2</sub>. Ideally, RI should be upgraded by dissolved O2 in the presence of NH4Cl from 42 % to approximately 94 % TiO<sub>2</sub>. However, the oxygen leaching of Fe<sup>0</sup> suffers from a serious limitations imposed by the sluggish mass transport rates of low soluble oxygen ( $\leq 8$  mg/L) to the interface Fe<sup>0</sup>/H<sub>2</sub>O. This physical barrier of iron corrosion products (oxide-layer or oxide-film) characterizes the similarity between the leaching process in the Becher process and contaminant removal in Fe<sup>0</sup>/H<sub>2</sub>O systems.

## Mechanism of iron removal in the Becher process

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Aqueous iron oxidation (corrosion) by dissolved oxygen (O<sub>2</sub>) has been extensively described in the literature, including the aeration step in the *Becher process* for the manufacture of synthetic rutile [14,15,18,19,22]. The overall reaction can be represented as:

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$$2 \text{ Fe}^0 + 3/2 \text{ O}_2 + \text{x H}_2\text{O} \Rightarrow \text{Fe}_2\text{O}_3.\text{xH}_2\text{O}$$
 [1]

The primary process is essentially an electrochemical reaction made up of (i) an anodic  $Fe^0$  oxidation (Eq. 2), and a cathodic  $O_2$  reduction (Eq. 3).  $Fe^{2+}$  ions from Eq. 2 are (at least partially) further oxidized by dissolved  $O_2$  to  $Fe^{3+}$  ions (Eq. 4), which are subsequently hydrolysed and precipitated as  $Fe_2O_3.(x-6)H_2O$  (Eq. 5).

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$$2 \text{ Fe}^0 \Rightarrow 2 \text{ Fe}^{2+} + 4 \text{ e}^{-}$$
 [2]

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$$O_2 + 2 H_2 O + 4 e^- \Rightarrow 4 HO^-$$
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$$4 \operatorname{Fe}^{2+} + \operatorname{O}_2 + 4 \operatorname{H}^+ \Rightarrow 4 \operatorname{Fe}^{3+} + 2 \operatorname{H}_2 \operatorname{O}$$
 [4]

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$$2 \text{ Fe}^{3+} + x \text{ H}_2\text{O} \Rightarrow \text{Fe}_2\text{O}_3.(x-6)\text{H}_2\text{O} + 6 \text{ H}^+$$
 [5]

- 82 It was established that the iron oxide products varied with time, yielding an iron oxide
- 83 mixture of unknown composition (FeOOH,  $Fe_2O_3$ ,  $Fe_3O_4$ ) [15,17].
- 84 A survey of the reactions above reveals three major features: (i) dissolved oxygen is consumed by both heterogeneous (electrochemical Fe<sup>0</sup> dissolution) and homogenous (Fe<sup>2+</sup> 85 oxidation) reactions; (ii) additional solid phases are formed (FeOOH, Fe<sub>2</sub>O<sub>3</sub>, Fe<sub>3</sub>O<sub>4</sub>) during 86 the process leading to a "three"-phase (Fe<sup>0</sup>-oxides-H<sub>2</sub>O) reaction system; (iii) in-situ formed 87 solid phases influence the rates of O<sub>2</sub> mass transfer to the Fe<sup>0</sup> surface. Clearly, iron oxides as 88 inert solid phase influence O2 mass transfer and thus the kinetics of heterogeneous reactions at 89 the interface Fe<sup>0</sup>/H<sub>2</sub>O. The impact of iron oxide will increase with increasing reaction time 90 since the amount of oxides is increasing. Accordingly, the heterogeneous reactions at the Fe<sup>0</sup> 91 92 surface can be divided into two phases: (i) an initial constant rate phase, followed by (ii) a

second phase, in which a steep decrease in rate is expected and was observed [15].

## Becher process versus contaminant removal in Fe<sup>0</sup>/H<sub>2</sub>O systems

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In the *Becher process* iron removal from a matrix of reduced ilmenite (RI - 58 % Fe<sup>0</sup> and 42 % TiO<sub>2</sub>) by oxidation with dissolved O<sub>2</sub> is the goal. The oxidation is performed in the presence of NH<sub>4</sub>Cl to limit iron oxide precipitation in the vicinity of RI (see below). Nevertheless, iron oxides precipitate and the goal of a matrix with 94 % TiO<sub>2</sub> is sometimes difficult to achieve [19,22]. Generated iron oxides are regarded as inert micro-particles influencing mass transfer between RI and dissolved O<sub>2</sub>. To accelerate mass transfer and thus maximize O<sub>2</sub> leaching in the *Becher process*, RI might be ground to small particle sizes and/or intensive solution mixing (stirring, agitating) yielding particle suspension might be applied.

In permeable reactive barriers, an elemental iron material (usually > 80 % Fe) is used; Fe<sup>0</sup> oxidation is ideally induced by the oxidized form of a contaminant. No external addition of chemical to limit iron oxide precipitation is suitable. The aquifer might contain chloride ions (Cl<sup>-</sup>) known for their destabilizing effects on the formation of oxide layers on Fe<sup>0</sup>. With regards to the objective (contaminant removal), Fe<sup>0</sup> oxidation by dissolved O<sub>2</sub> should be considered a disturbing factor yielding iron oxides that should act as physical barrier for contaminant transport to the Fe<sup>0</sup>/H<sub>2</sub>O interface (as in the Becher process). This is the fundamental role of an oxide-layer (oxide-film), which should have been considered as the rule. The exceptions of conducting oxide-layers or lattice Fe<sup>II</sup> surface sites [13,23,24] could be treated separately. This has not been the case, showing that lessons from the Becher process are very helpful for the development of the Fe<sup>0</sup> technology. Contrary to the Becher process, acceleration of mass transfer of contaminant to the Fe<sup>0</sup>/H<sub>2</sub>O interface should be limited to conditions of mass-transfer control [11,12]. Under these conditions material suspension should not occur and the formation of the oxide layer on Fe<sup>0</sup> should not be significantly disturbed. The most important features common to the Becher process and contaminant removal in remediation Fe<sup>0</sup>/H<sub>2</sub>O systems are: (i) iron dissolution and iron precipitation yielding a complex oxide mixture of unknown composition, (ii) Fe<sup>0</sup> and Fe<sup>II</sup> are potential reducing agents, and (iii) the weight fraction of iron oxide particles would increase from zero at the beginning to higher proportions depending on the reaction progress. The net result for both processes is a multi-solid reaction system involving two types of solids of widely differing size (RI/Fe<sup>0</sup> and iron oxides). The large changes in the solid composition during the reaction can be expected to influence the mass transfer of species to the Fe<sup>0</sup>/H<sub>2</sub>O interface and thereby play a significant role in the determination of reaction rates. The major difference between both processes is that iron oxides can be regarded as inert in the Becher process but are contaminant adsorbents in Fe<sup>0</sup>/H<sub>2</sub>O systems. Furthermore, during their

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polymerisation and precipitation, iron (hydr)oxides may sequester contaminants in their matrix (so-called co-precipitation) [11,12]. Therefore, contaminants can be immobilized by oxide-layers on Fe<sup>0</sup> regardless of any chemical or electrochemical redox reaction, possibly involving Fe<sup>0</sup>. Finally, oxides in Fe<sup>0</sup>/H<sub>2</sub>O systems are good adsorbents for Fe<sup>II</sup> from Fe<sup>0</sup> oxidation yielding adsorbed Fe<sup>II</sup> (structural Fe<sup>II</sup>). Structural Fe<sup>II</sup> is a very strong reducing agent for inorganics and organics [25]. Table 1 summarises some relevant processes involved in remediation Fe<sup>0</sup>/H<sub>2</sub>O systems and in the *Becher process*.

It should be acknowledged that increasing attention has been paid to the role of oxide layer on Fe<sup>0</sup> in the mechanisms of contaminant reductive transformation in the literature [8,10,13,24,26-29,35]. However, these experiments were mostly performed to gain information on the nature of surface coatings and their influence on kinetics and reaction pathways of the reductive transformation of water pollutants. While these experiments show the role of oxide layers (iron oxides/hydroxides) on Fe<sup>0</sup> as possible electron mediators, the

Becher process indicates that oxide layers fundamentally act as physical barriers. Recent

works [11,12] demonstrated that contaminants are primarily removed by adsorption and co-

precipitation within the oxide layers and are stable as long as the iron oxide is not dissolved.

### **Lessons from the** *Becher process***:**

Modelling is now a fundamental step in science, since it allows a better understanding of investigated systems. Modelling can be used to support experimental design, operation control, experiment planning, experiment optimization, experiment scaling, and system or results analysis [30]. Upon adequately representing a system, one might be able to scale-up, design, optimize, or control a system. To be efficient, each model needs (i) relevant parameters, (ii) relevant data, and (iii) expertise of the modeller in terms of adequate computational methods. Models for the adequate understanding of remediation  $Fe^0/H_2O$  systems could benefit from expertise from the *Becher process* if they considered the abovementioned differences (see also table 1). In particular, the interplay between mass transfer and

kinetics encountered in the Becher process will help to better plan experiments as well as improve the interpretation of results from extensive studies on the kinetics of contaminant removal in Fe<sup>0</sup>/H<sub>2</sub>O systems. Similar to the *Becher process*, this involves both Fe<sup>0</sup> and Fe<sup>II</sup> oxidation by relevant dissolved species (including contaminants and oxygen). Several different types of models have been developed with regards to the *Becher process* [17,18]. These models consider various aspects of the operation such as O<sub>2</sub> mass transfer and the presence of solid, liquid and gaseous phases. For example, the model developed by Geetha and Surender [18] considered the *Becher process* as a multi-phase system involving a complex interplay between mass transfer and chemical reactions. This model can be further developed to meet the specifics of remediation Fe<sup>0</sup>/H<sub>2</sub>O systems. In particular the following aspects need to be considered: (i) the fact that iron dissolution should be coupled to contaminant removal, (ii) the limited availability of oxygen, which is a disturbing factor (concurrent for Fe<sup>0</sup> oxidation), (iii) the fact that iron oxides are not inert particles on mass transfer rates but effective contaminant removing agents (adsorption and co-precipitation), and (iv) the specific interactions of individual contaminants with the oxide layer on Fe<sup>0</sup>. It can be emphasized that the overall process of contaminant removal under field conditions will be controlled by an internal diffusion through the oxide layer, since almost all contaminants are larger in size than molecular oxygen (the oxidant of the *Becher process*). Another important result from the *Becher process* is highlighted by the work of Farrow et al. [21] who compared the role of sodium chloride (NaCl) and ammonium chloride (NH<sub>4</sub>Cl) on the process of selective iron leaching. It is well-known that the chloride ion plays an important role in the breakdown of oxide layers and is therefore a desirable constituent in any electrolyte solution used for the promotion of iron corrosion [31,32]. To elucidate the role of ammonium ions in the leaching reaction, parallel experiments with 0.1 M of each chloride were carried out at 70°C. In both cases the dissolved iron concentration and the bulk solution pH were monitored during the course of the reaction. The results showed that the ammonium

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chloride plays three distinct and important roles in the selective leaching [21]: (i) NH<sub>4</sub><sup>+</sup> acts as a buffer for hydroxyl ions (OH<sup>-</sup>) and prevents excessively high local pH values, which might otherwise cause precipitation of iron (II) hydroxide before the iron (II) ions could diffuse from the rutile matrix; (ii) the ammonia (NH<sub>3</sub>) formed as a result of this buffering reaction complexes iron (II) ions until they also have moved away from the regions of high pH, thus preventing the precipitation of iron (II) hydroxide in the pores of the rutile; (iii) the chloride ion helps to break down any passive films which might form during aeration. The large difference in the leaching rates for NaCl and NH<sub>4</sub>Cl could be explained using pH data, which indicate that most of the reactions in NH<sub>4</sub>Cl takes place at a pH of about 4, and most of the reactions in NaCl takes place at a pH between 9 and 10. These results show clearly that in investigating process relevant for remediation Fe<sup>0</sup>/H<sub>2</sub>O systems while avoiding iron oxide precipitation, it will be more advantageous to work with NH<sub>4</sub>Cl than with strong chelating agents like ethylenediaminetetraacetic acid [1,33] of larger molecular size and little buffering capacity. A further relevant result from the Becher process is obtained by Bruckard et al. [22] on the optimisation of selective iron leaching using the anthraquinone-based redox catalysts while guaranteeing the formation of the preferred magnetite as the reaction product. This result could be very important for enhancing long-term reactivity of Fe<sup>0</sup>/H<sub>2</sub>O systems. The redox catalysts are effective for three main reasons [22, 34]: (i) their electrochemical reactions are fast and reversible, (ii) their redox potentials are located between those of the Fe<sup>II</sup>/Fe<sup>0</sup> and H<sub>2</sub>O/O<sub>2</sub> couples such that their oxidised forms (the anthraquinones) oxidise metallic iron and their reduced forms (the semi-quinones or hydroquinones) are oxidised by oxygen, and (iii) both the reduced and oxidised forms of the catalysts are stable in solution. A modest enhancement of contaminant reduction by Fe<sup>0</sup> in the presence of several quinonoid model compounds (e.g. juglone, lawsone, and anthraquinone disulfonate) was also reported by Tratneyk et al. [24,35] in remediation Fe<sup>0</sup>/H<sub>2</sub>O systems. However, the observed effects were

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not expected to be significant in field applications. Yet, if quinonoid model compounds effectively favour the formation of electronic conductive magnetite (Fe<sub>3</sub>O<sub>4</sub>) as reported by Bruckard et al. [22], their effects for long-term reactivity of Fe<sup>0</sup> walls should be significant. More research under conditions pertinent to remediation Fe<sup>0</sup>/H<sub>2</sub>O systems is needed to clarify this issue.

### **Concluding remarks**

While the complexity of mass-transfer in remediation Fe<sup>0</sup>/H<sub>2</sub>O systems is yet to be properly assessed, the *Becher process* for the manufacture of synthetic rutile has already dealt with the prediction of solid-liquid mass transfer in systems in with particles of various densities and sizes are available [15,17-19]. The results achieved in developing the *Becher process* could be very useful to investigations regarding processes in Fe<sup>0</sup> walls. In particular, the interplay between mass transfer and kinetics reported in the *Becher process* could help in improving the interpretation of data from extensive investigations on the kinetics of contaminant removal in Fe<sup>0</sup>/H<sub>2</sub>O systems. For this purpose, specifics of the Fe<sup>0</sup>/H<sub>2</sub>O system (limited O<sub>2</sub> availability, reactivity of the oxide layer, dynamic nature of the film formation and transformation) should to be properly considered and data should be obtained under relevant experimental conditions [11,12]. Furthermore, electron mediators have been unambiguously described in the *Becher process* [22]. Electron mediators facilitate electron transport from Fe<sup>0</sup> to reducible species (oxygen, contaminants), they can undergo many successive reduction/oxidation cycles in the process and act as catalysts. Catalytic effects of quinonoid model compounds in remediation Fe<sup>0</sup>/H<sub>2</sub>O systems should be further investigated.

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- 235 References
- 236 [1] L.J. Matheson, P.G. Tratnyek, Reductive dehalogenation of chlorinated methanes by iron
- 237 metal, Environ. Sci. Technol. 28 (1994), 2045-2053.
- 238 [2] R. Focht, J. Vogan, S. O'Hannesin, Field application of reactive iron walls for in-situ
- degradation of volatile organic compounds in groundwater. Remediation 6 (1996), 81-94.
- 240 [3] S.F. O'Hannesin, R.W. Gillham, Long-term performance of an in situ "iron wall" for
- 241 remediation of VOCs. Ground Water 36 (1998), 164-170.
- 242 [4] J.L. Jambor, M. Raudsepp, K. Mountjoy, Mineralogy of permeable reactive barriers for
- 243 the attenuation of subsurface contaminants. Can. Miner. 43 (2005), 2117-2140.
- 244 [5] A.D. Henderson, A.H. Demond, Long-term performance of zero-valent iron permeable
- reactive barriers: a critical review. Environ. Eng. Sci. 24 (2007), 401-423.
- 246 [6] R.L. Johnson, R.B. Thoms, R.O'B. Johnson, J.T. Nurmi, P.G. Tratnyek, Mineral
- 247 precipitation upgradient from a zero-valent iron permeable reactive barrier. Ground Water
- 248 Monit. Remed. 28 (2008), 56-64.
- [7] R.L. Johnson, R.B. Thoms, R.O'B. Johnson, T. Krug, Field evidence for flow reduction
- 250 through a zero-valent iron permeable reactive barrier. Ground Water Monit. Remed. 28
- 251 (2008), 47-55.
- 252 [8] E.J. Weber, Iron-mediated reductive transformations: investigation of reaction
- 253 mechanism. Environ. Sci. Technol. 30 (1996), 716-719.
- 254 [9] B.K. Lavine, G. Auslander, J. Ritter, Polarographic studies of zero valent iron as a
- reductant for remediation of nitroaromatics in the environment. Microchem. J., 70 (2001), 69-
- 256 83.

- 257 [10] J.A. Mielczarski, G.M. Atenas, E. Mielczarski, Role of iron surface oxidation layers in
- decomposition of azo-dye water pollutants in weak acidic solutions. Appl. Catal. B: Environ.
- 259 56 (2005), 289-303.
- 260 [11] C. Noubactep, Processes of contaminant removal in "Fe<sup>0</sup>–H<sub>2</sub>O" systems revisited: The
- importance of co-precipitation. Open Environ. J. 1 (2007), 9-13.
- 262 [12] C. Noubactep, A critical review on the mechanism of contaminant removal in Fe<sup>0</sup>–H<sub>2</sub>O
- 263 systems. Environ. Technol. 29 (2008), 909-920.
- 264 [13] M.M. Scherer, B.A. Balko, P.G. Tratnyek, The role of oxides in reduction reactions at
- 265 the metal-water interface, in: D. Sparks, T. Grundl (Eds.), Kinetics and Mechanism of
- Reactions at the Mineral/Water Interface, D. Sparks, T. Grundl, Eds; American Chemical
- 267 Society: Washington, DC. 1998, pp. 301-322.
- 268 [14] R.G. Becher, R.G. Canning, B.A. Goodheart, S. Uusna, A new process for upgrading
- ilmenitic sands. Proc. Aust. Inst. Min. Metall. 21 (1965), 21-44.
- 270 [15] K.S. Geetha, G.D. Surender, Solid-liquid mass transfer in the presence of micro-particles
- during dissolution of iron in a mechanically agitated contactor. Hydrometallurgy 36 (1994),
- 272 231-246.
- 273 [16] Y. Chen, T. Hwang, M. Marsh, J.S. Williams, Mechanically activated carbothermic
- 274 reduction of ilmenite. Metal. Mater. Trans. A 28 (1997), 1115-1121.
- 275 [17] K.S. Geetha, G.D. Surender, Modelling of ammoniacal oxygen leaching of metallic iron
- in a stirred slurry reactor. Hydrometallurgy 44 (1997), 213-230.
- 277 [18] K.S. Geetha, G.D. Surender, Experimental and modelling studies on the aeration
- 278 leaching process for metallic iron removal in the manufacture of synthetic rutile.
- 279 Hydrometallurgy 56 (2000), 41-62.
- 280 [19] K.S. Geetha, G.D. Surender, Intensification of iron removal rate during oxygen leaching
- through gas-liquid mass-transfer enhancement. Metall. Mater. Trans. B32 (2001), 961-963.

- 282 [20] K.M. Blyth, M.I. Ogden, D.N. Phillips, D. Pritchard, W. van Bronswijk, Reduction of
- 283 Ilmenite with Charcoal. J. Chem. Educ. 82 (2005), 456-459.
- 284 [21] J.B. Farrow, I.M. Ritchie, P. Mangano, The reaction between reduced ilmenite and
- 285 oxygen in ammonium chloride solutions. Hydrometallurgy 18 (1987), 21-38.
- 286 [22] W.J. Bruckard, C. Calle, S. Fletcher, M.D. Horne, G.J. Sparrow, A.J. Urban, The
- application of anthraquinone redox catalysts for accelerating the aeration step in the becher
- 288 process. Hydrometallurgy 73 (2004), 111-121.
- 289 [23] M.M. Scherer, S. Richter, R.L. Valentine, P.J.J. Alvarez, Chemistry and microbiology of
- 290 permeable reactive barriers for in situ groundwater clean up. Rev. Environ. Sci. Technol. 30
- 291 (2000), 363-411.
- 292 [24] P.G. Tratnyek, M.M. Scherer, B. Deng, S. Hu, Effects of natural organic matter,
- anthropogenic surfactants, and model quinones on the reduction of contaminants by zero-
- 294 valent iron. Wat. Res. 35 (2001), 4435-4443.
- 295 [25] A.F. White, M.L. Peterson, Reduction of aqueous transition metal species on the surfaces
- of Fe(II)-containing oxides. Geochim. Cosmochim. Acta 60 (1996), 3799-814.
- 297 [26] M.S. Odziemkowski, T.T. Schuhmacher, R.W. Gillham, E.J. Reardon, Mechanism of
- 298 oxide film formation on iron in simulating groundwater solutions: raman spectroscopic
- 299 studies, Corr. Sci. 40 (1998), 371-389.
- 300 [27] P.M.L. Bonin, M.S. Odziemkowski, R.W. Gillham, Electrochemical and Raman
- 301 spectroscopic studies of the influence of chlorinated solvents on the corrosion behaviour of
- iron in borate buffer and in simulated groundwater. Corros. Sci. 40 (1998) 1921-1939.
- 303 [28] P.M.L. Bonin, W. Jedral, M.S. Odziemkowski, R.W. Gillham, Electrochemical and
- Raman spectroscopic studies of the influence of chlorinated solvents on the corrosion
- behaviour of iron in borate buffer and in simulated groundwater. Corros. Sci. 42 (2000) 1921-
- 306 1939.

- 307 [29] M.S. Odziemkowski, R.P. Simpraga, Distribution of oxides on iron materials used for
- 308 remediation of organic groundwater contaminants Implications for hydrogen evolution
- 309 reactions. Can. J. Chem./Rev. Can. Chim. 82 (2004), 1495-1506.
- 310 [30] M.E. Mellado, L.A. Cisternas, E.D. Gálvez, An analytical model approach to heap
- 311 leaching. Hydrometallurgy 95 (2009), 33-38.
- 312 [31] Speller F., Corrosion Causes and Prevention, McGraw-Hill, New York, (1951) 686 pp.
- 313 [32] J.S. Kim, P.J. Shea, J.E. Yang, J.-E. Kim, Halide salts accelerate degradation of high
- explosives by zerovalent iron. Environ. Pollut. 147 (2007), 634-641.
- 315 [33] J.-L. Chen,, S.R. Al-Abed, J.A. Ryan, Z. Li, Effects of pH on dechlorination of
- trichloroethylene by zero-valent iron. J. Hazard. Mater. B83 (2001) 243-254.
- 317 [34] R.P. Schwarzenbach, R.Stierli, K.Lanz, J. Zeyer, Quinone and iron porphyrin mediated
- 318 reduction of nitroaromatic compounds in homogeneous aqueous solution. Environ. Sci.
- 319 Technol. 24 (1990), 1566-1574.
- 320 [35] T.L. Johnson, W. Fish, Y.A. Gorby, P.G. Tratnyek, Degradation of carbon tetrachloride
- 321 by iron metal: Complexation effects on the oxide surface. J. Cont. Hydrol. 29 (1998), 379-
- 322 398.

**Table 1:** Overview of some relevant processes involved in remediation Fe<sup>0</sup>/H<sub>2</sub>O systems and in the *Becher Process*. In both cases the anodic reaction is the electrochemical dissolution of Fe<sup>0</sup> which is followed by the chemical precipitation of Fe<sup>II</sup>/Fe<sup>III</sup> oxyhydroxides (oxide layer). Irrespective of the cathodic reactants (O<sub>2</sub>, contaminants), the rate of aqueous iron corrosion is proportional to the total amount of reducible species transferred to the Fe<sup>0</sup> surface. In the *Becher Process*, enhanced O<sub>2</sub> mass-transfer rate can be achieved by various mixing operations and the extent of Fe<sup>0</sup> dissolution increased for example by using ammonium chloride. In groundwater remediation species transport to Fe<sup>0</sup> depends on natural hydrodynamic conditions.

Process	<b>Groundwater treatment</b>	Becher process
<b>Objective</b>	Contaminant removal	Upgrading of reduced ilmenite (42 % TiO <sub>2</sub> )
		to synthethic rutile (94 % TiO <sub>2</sub> )
<b>Used material</b>	Fe-based material (up to 99.0 % Fe)	Reduced ilmenite (approx. 58 % Fe)
Method	Oxidation of Fe <sup>0</sup> by contaminants	Selective oxidation of Fe <sup>0</sup> by O <sub>2</sub>
Role of Fe <sup>II</sup>	suitable contaminant reducing agent	O <sub>2</sub> scavenger within the oxide layer
	within the oxide layer	(non-relevant process – side reaction)
Nature	Environmental process	Industrial process
	(controlled exclusively by	(O <sub>2</sub> leaching process can be intensified
	environmental conditions)	e.g., by the use of ferrous chloride)
Conditions	6.0 < pH < 9.5	pH > 6.0
	Temperature lower than 20 °C	Temperature: ambient to 130 °C
	limited O <sub>2</sub> availabiliy (< 2 mg/L)	O <sub>2</sub> available (ideally 8 mg/L)
	no iron oxyhydroxides at $t = 0$	no iron oxyhydroxides at $t = 0$
	various iron oxyhydroxides at $t > 0$	various iron oxyhydroxides at $t > 0$
Limitations	Fe-oxides limit mass transport rates	Fe-oxides limit mass transport rates
(passivation)	of species (incl. O <sub>2</sub> ) to the Fe <sup>0</sup>	of oxygen to the Fe <sup>0</sup> surface
Remedy to	Use of bimetallics and nanoparticles use of ammonium chloride as a buffering	
passivation	Use of ultrasound.	and complexing agent (initial $pH = 4.5$ )